



Evaluating the
climate and air
quality impacts of
short-lived pollutants

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Evaluating the climate and air quality impacts of short-lived pollutants

A. Stohl¹, B. Aamaas², M. Amann³, L. H. Baker⁴, N. Bellouin⁴, T. K. Berntsen², O. Boucher⁵, R. Cherian⁶, W. Collins^{4,7}, N. Daskalakis^{8,9}, M. Dusinska¹, S. Eckhardt¹, J. S. Fuglestedt², M. Harju¹, C. Heyes³, Ø. Hodnebrog², J. Hao¹⁰, U. Im^{8,*}, M. Kanakidou^{8,9}, Z. Klimont³, K. Kupiainen³, K. S. Law¹¹, M. T. Lund², R. Maas¹², C. R. MacIntosh⁴, G. Myhre², S. Myriokefalitakis^{8,9}, D. Olivie¹³, J. Quaas⁶, B. Quennehen¹¹, J.-C. Raut¹¹, S. T. Rumbold⁷, B. H. Samset², M. Schulz¹³, Ø. Seland¹³, K. P. Shine⁴, R. B. Skeie², S. Wang¹⁰, K. E. Yttri¹, and T. Zhu¹⁴

¹NILU – Norwegian Institute for Air Research, Kjeller, Norway

²Center for International Climate and Environmental Research – Oslo (CICERO), Oslo, Norway

³International Institute for Applied Systems Analysis (IIASA), Laxenburg, Austria

⁴Department of Meteorology, University of Reading, Reading, United Kingdom

⁵LATMOS, Université Pierre et Marie Curie (UPMC)/CNRS, Paris, France

⁶Institute for Meteorology, Universität Leipzig, Germany

⁷Met Office Hadley Centre, Exeter, United Kingdom

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⁸Environmental Chemical Processes Laboratory, Department of Chemistry, University of Crete, Heraklion, Crete, Greece

⁹FORTH, ICE-HT, Platani, Patras, Greece

¹⁰School of Environment, Tsinghua University, Beijing, China

¹¹Sorbonne Universités, UPMC Univ. Paris 06, Université Versailles St-Quentin, CNRS/INSU, LATMOS-IPSL, UMR 8190, Paris, France

¹²RIVM – National Institute for Public Health and the Environment, Bilthoven, the Netherlands

¹³Norwegian Meteorological Institute, Oslo, Norway

¹⁴State Key Laboratory for Environmental Simulation and Pollution Control, College of Environmental Sciences and Engineering, Peking University, Beijing, China

* now at: Aarhus University, Department of Environmental Science, Frederiksborgvej 399, 4000 Roskilde, Denmark

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Correspondence to: A. Stohl (ast@nilu.no)

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Europe, where the surface warming was reduced by about 0.3 K and precipitation rates were increased by about 15 (6–21) mm yr⁻¹ (more than 4 % of total precipitation) from spring to autumn. Thus, the mitigation could help to alleviate expected future drought and water shortages in the Mediterranean area. We also report other important results of the ECLIPSE project.

1 Introduction

The United Nations Framework Convention on Climate Change (UNFCCC) requires climate policies to “be cost-effective so as to ensure global benefits at the lowest possible cost” and that “policies and measures should ... be comprehensive ... [and] ... cover all relevant sources, sinks and reservoirs”. This was made operational by the Kyoto Protocol, which sets limits on emissions of six different greenhouse gases (GHGs), or groups of GHGs – carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O), perfluorocarbons (PFCs), hydrofluorocarbons (HFCs) and sulfur hexafluoride (SF₆). Collectively these are often known as “the Kyoto gases” or the “Kyoto basket”¹. CO₂ is the most important anthropogenic driver of global warming, with additional significant contributions from CH₄ and N₂O. However, other anthropogenic emissions capable of causing climate change are not covered by the Kyoto Protocol. Some are covered by other protocols, e.g. emissions of chlorofluorocarbons (CFCs) and hydrochlorofluorocarbons (HCFCs) are regulated by the Montreal Protocol, because of their role in stratospheric ozone (O₃) depletion. But there are others, notably several short-lived

¹Note that, formally, only species given values of Global Warming Potentials (GWP) in IPCC’s Second Assessment Report were controlled during the first commitment period (2008–2012) of the Kyoto Protocol. The second commitment period (2013–2020), via the Doha Amendment, also includes NF₃ in the list of greenhouse gases, and uses GWP values from the IPCC’s Fourth Assessment Report. The Doha Amendment is currently not in force, as it awaits ratification by a sufficient number of parties.

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components that give strong contributions to climate change that are excluded from existing climate agreements.

In the present study we investigate climate and air quality impacts of the emissions of CH₄, which has a lifetime of about 9 ± 1 years (Prather et al., 2012) and a number of much shorter-lived components (atmospheric lifetimes of months or less) which directly or indirectly (via formation of other short-lived species) influence the climate (Myhre et al., 2013):

- Methane is a greenhouse gas roughly 26 times stronger than CO₂ on a per molecule basis at current concentrations. It is relatively well-mixed in the atmosphere and has both natural and anthropogenic sources. It is also a precursor of O₃ and stratospheric water vapour.
- Black carbon (BC, also commonly known as soot), a product of incomplete combustion of fossil fuels and biomass, affects climate via several mechanisms (Bond et al., 2013). It causes warming through absorption of sunlight and by reducing surface albedo when deposited on snow. BC also affects clouds, with a consequent (but highly uncertain) impact on their distribution and radiative properties (Boucher et al., 2013).
- Tropospheric O₃ is a greenhouse gas produced by chemical reactions from the emissions of the precursors CH₄, carbon monoxide (CO), non-CH₄ volatile organic compounds (NMVOCs) and nitrogen oxides (NO_x). Emissions of these same precursors also impact on hydroxyl radical (OH) concentrations with further impacts especially on CH₄.
- Several components have cooling effects on climate, mainly sulphate aerosol formed from sulphur dioxide (SO₂) and ammonia (NH₃), nitrate aerosol formed from NO_x and NH₃, and organic aerosol (OA) which can be directly emitted or formed from gas-to-particle conversion of NMVOCs. They cause a direct cooling by scattering solar radiation and alter the radiative properties of clouds, very likely leading to further cooling.

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We refer to these substances as short-lived climate pollutants (SLCPs) as they also have detrimental impacts on air quality, directly or via formation of secondary pollutants (Kirtman et al., 2013). Notice that we include the precursors of O₃ and secondary aerosols in our definition of SLCPs. We also include CH₄ in our study even though it is included in the Kyoto Protocol, because of its relatively short lifetime compared to that of CO₂ and its importance for air quality via the formation of O₃. We do not include HFCs in our definition of SLCPs, as they have no significant impact on air quality and can be regulated from a climate policy perspective alone. For SLCPs, on the other hand, cost-effective environmental policy measures should be designed such that they optimize both the climate and air quality responses (Schmale et al., 2014). In some instances, control of the emissions of a species is expected to reduce future warming and improve air quality at the same time – a “win-win” situation (Anenberg et al., 2012); in others, the control of emissions may be conflicting, in the sense that it could increase warming while improving air quality (or vice versa) – in this case, emission control involves a “trade-off” between the impacts.

The net climate impact since pre-industrial times of all short-lived components other than CH₄ together is very likely to be cooling due primarily to sulphate aerosols (Myhre et al., 2013). Whilst SLCP reductions are clearly beneficial for air quality, elimination of all current non-CH₄ SLCP emissions would thus very likely lead to extra warming. Nevertheless, targeted emission reductions of selected SLCPs which cause warming (either directly or via formation of secondary species) have the potential to reduce global warming on a short timescale, as well as improving air quality. They may also reduce the rate of warming (Myhre et al., 2011; Shindell et al., 2012) that is important, for example, for the adaptation of ecosystems to climate change (as recognized by UNFCCC Art. 2) and is expected to accelerate in the near future (Smith et al., 2015). Reducing these selected SLCP emissions might be effective to help avoid (or at least delay) certain undesired impacts of climate change (e.g., rapid sea ice loss in the Arctic; Quinn et al., 2008). At least, optimized SLCP emission reductions could help to reduce

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Second, the climate impact of short-lived components, even when averaged globally, can depend strongly on location and time (e.g., summer vs. winter) of emissions (Fuglestvedt et al., 1999; Wild et al., 2001; Berntsen et al., 2005, 2006; Koch et al., 2007; Naik et al., 2005; Reddy and Boucher, 2007; Shindell and Faluvegi, 2009). For well-mixed gases (e.g. Kyoto gases), a single globally-valid value of Global Warming Potential (GWP – see Sect. 1.2 for more details) can be calculated for a chosen time horizon, and then used to give the so-called “CO₂-equivalent” emissions of a gas. By contrast, for the non-CH₄ SLCPs, the GWP depends significantly on when and where the emission occurs. Not only does this complicate the calculation of GWPs, it also introduces an additional dimension into the framing of climate policy. For instance, the importance of location for BC emissions has received much attention in this context (Ramanathan and Carmichael, 2008; Shindell and Faluvegi, 2009).

Third, inhomogeneity in the climate response to RF is important for SLCPs. The geographical pattern of RF due to the non-CH₄ SLCPs is generally concentrated close to the source of emission, and hence is quite distinct from the global-scale forcing due to the Kyoto gases. The extent to which these heterogeneous forcing patterns will trigger different climate responses compared to well-mixed gases is an unresolved scientific issue, even though the climate response generally occurs on larger spatial scales (but mainly in the hemisphere where the forcing takes place; Joshi et al., 2003; Shindell et al., 2010) than the forcing itself. One example of the issue of inhomogeneity of response concerns the effects of absorption of solar radiation by BC in the Arctic atmosphere. Flanner (2013) has shown that in the Arctic BC located at low altitudes causes a strong local surface warming, but BC located at higher altitudes causes a surface cooling, which is due to the reduced solar radiation reaching the surface. Another important example is emissions of NO_x as these lead to a shorter-lived (and hence more localised) positive RF due to increases in O₃ and a longer-lived (and hence more global) negative RF due to the increased rate of destruction of CH₄. This means that metrics based on global-mean quantities may be poorly representative of the local im-

by the UNECE Convention on Long-Range Transboundary Air Pollution (CLRTAP) and its protocols including the Gothenburg Protocol and its amendments.

The International Agency for Research on Cancer classified outdoor air pollution as carcinogenic to humans with sufficient evidence that it causes lung cancer. A positive association with an increased risk of bladder cancer was also demonstrated. It has been estimated that air pollution caused 223 000 deaths from lung cancer worldwide in 2010 (Unger et al., 2010). Air quality guidelines for various substances published by different agencies are listed in Table 1.

Ozone and PM are the most problematic air pollutants with regard to effects on human health (EEA, 2013). Ozone can, through impairment of lung function, lead to premature deaths and increased hospitalization (West et al., 2006). PM was classified as carcinogenic to humans (IARC, 2015; Grosse, 2013). It is estimated, for instance, that an increase of $10 \mu\text{g m}^{-3}$ in the concentrations of PM_{10} (PM with diameter smaller than $10 \mu\text{m}$) will increase cardiopulmonary mortality by 9% (Pope et al., 1995). Different aerosol types are considered when assessing climate impacts, whereas air quality legislation is based on the concept of total mass concentrations of particulate matter – either as $\text{PM}_{2.5}$ or PM_{10} . It is, however, likely that human health impacts also depend on PM composition. For instance, according to the World Health Organization (WHO), epidemiological evidence indicates an association of daily variation in BC concentrations with short and long term adverse health effects such as all-cause and cardiovascular mortality, and cardiopulmonary hospital admissions. Additionally, BC was classified as possibly carcinogenic to humans (Group 2B) (WHO, 2012). However, concentration-response functions for individual PM components still need to be established. Thus, neither BC nor ultrafine particles are currently covered specifically by EU guidelines (WHO, 2013).

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a beneficial air quality impact were then evaluated according to their expected climate benefit. Notice here that emission measures typically affect several SLCP species. For every mitigation measure, the emission reduction of every SLCP species was therefore weighted with the chosen climate metric and summed over all emitted SLCP species.

5 Finally, all measures with beneficial air quality and climate impacts were collected in a basket defining the SLCP mitigation scenario (see Sect. 3.5).

The metrics, however, cannot fully quantify the climate response, due to the underlying simplifying assumptions, including linearity, the need to specify particular time horizons and, most importantly, the focus on one single aspect of climate change (global mean temperature for the chosen GTP metric). Therefore, a second research path (the inner part of the spiral in Fig. 1) was taken to determine the climate response for a set of emission reductions for individual SLCP species, using a small ensemble of four advanced Earth system models (ESMs). Furthermore, transient climate ensemble simulations with these ESMs were run for the baseline and emission mitigation scenarios, to calculate the transient climate and air quality impacts of the mitigation scenario (see Sect. 3.6). A comparison between the climate impacts expected from the metrics and those calculated with the transient simulations (left part of the spiral in Fig. 1) closed the loop between the first and the second research path and allowed the evaluation of the consistency of both approaches (see Sect. 3.7).

20 3 Results

3.1 The ECLIPSE emissions

The ECLIPSE emission data set was created with the GAINS (Greenhouse gas – Air pollution Interactions and Synergies; <http://gains.iiasa.ac.at>) model (Amann et al., 2011), which provides emissions of long-lived greenhouse gases and shorter-lived species in a consistent framework. The GAINS model holds essential information about

25 sources of emissions, environmental policies, and mitigation opportunities for about

sion Database (GFED), version 3.1 (van der Werf et al., 2010) for the years 2008 and 2009 and held constant in simulations of future scenarios. Biogenic emissions originate from the MEGAN database (Guenther et al., 2012; <http://lar.wsu.edu/megan/>).

3.2 Model evaluation

Using the ECLIPSE version 4a CLE emissions, simulations were carried out with a range of models. In addition to the four ESMs used in ECLIPSE (HadGEM3, ECHAM6-HAM2, NorESM1-M and CESM1/CAM5.2; see Baker et al., 2015a for descriptions of these models), three CTMs and a Lagrangian particle dispersion model were used (see Table 2). All models were run for core periods in 2008 and 2009, when several aircraft campaigns took place in China and the Arctic, but most models simulated the full 2008–2009 period. Some models were also run for longer periods and were evaluated together with other models. For instance, in a comparison against aircraft measurements, Samset et al. (2014) found that the models systematically over-predict BC concentrations in the remote troposphere, especially at higher altitudes. They concluded that the BC lifetime in the models is too long. A follow-up study suggested that the best match to aircraft observations could be achieved with strongly increased BC emissions and decreased lifetimes (Hodnebrog et al., 2014). Daskalakis et al. (2015) derived changes in the local lifetime of BC up to 150 % associated with the use of different amounts and spatial distribution of fire emissions in the same chemistry transport model, demonstrating the dependence of BC lifetime on its emissions. Tsigaridis et al. (2014) found systematic underprediction of OA near the surface as well as a large model divergence in the middle and high troposphere. They attributed these discrepancies to missing or underestimated OA sources, the removal parameterisations as well as uncertainties in the temperature-dependent partitioning of secondary OA in the models. As a consequence of these studies, ECLIPSE models were improved in terms of emissions, secondary OA formation and removal parameterisations.

A detailed evaluation of the performance of the improved ECLIPSE models was made for Eastern Asia, using satellite, airborne and ground-based measurements of

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pollutant gases (CO, NO₂, O₃ and SO₂) and aerosols (Quennehen et al., 2015). For Eastern Asia in August–September 2008 (Fig. 3, left two columns), data were averaged over three urban and five rural sites. The models have difficulties reproducing the urban concentrations, due to their coarse resolution. However, surprisingly most models over-estimated the urban SO₂ mixing ratios. This could be related to power plant emissions that are actually occurring outside urban boundaries, being placed into the coarse urban model grid cells. For urban NO₂ models deviate less from observations, with both overestimates and underestimates, thus the model mean captures the observations. The measured concentrations of NO₂ and O₃ at the rural sites are matched relatively well (agreement within the range of the temporal distribution at individual sites) but SO₂ is generally overestimated there as well. The most severe problem at rural sites, however, is a systematic underestimation of CO mixing ratios, which was attributed to underestimated CO lifetimes in the models (Quennehen et al., 2015).

A similar comparison was made for Europe with background measurements taken from stations of the European Monitoring and Evaluation Programme (EMEP) (Fig. 3, right two columns for winter and summer). Overall, over Europe the ECLIPSE model mean captures the mean observations with the exception during summer for CO that is underestimated (as in Asia). Summertime O₃ is overestimated by many models at rural locations over Europe and Asia suggesting too much photochemical production downwind of emission regions.

Satellite-derived aerosol optical depth (AOD) measurements were reproduced quite well by the models over China and Europe (Fig. 4). Evaluation of individual aerosol components over Asia (Quennehen et al., 2015) shows an overestimation of the ECLIPSE model-mean surface BC in urban China in summer 2008, which is probably due to the short-term mitigation measures taken during the Olympic Games. Over Europe, ECLIPSE models satisfactorily simulate surface BC observations both in winter and summer (Fig. 4). However, problems were identified over India: Gadhavi et al. (2015) found that BC concentrations are strongly underestimated in Southern India even when aerosol removal processes in one model were completely switched off

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in the region. Furthermore, observed AOD values in Northern India are larger than simulated by all but two of the ECLIPSE models (Fig. 4). This suggests that the emissions of BC and precursors of other aerosols are underestimated for India in the ECLIPSE emission data set. This could be related to the rapid recent growth of emissions in India (Klimont et al., 2013), which may be underestimated in the inventories, as well as with problems capturing the true spatial distribution of emissions in India.

The Arctic was shown previously to be a particularly challenging region for aerosol model simulations (e.g., Shindell et al., 2008). An evaluation of the ECLIPSE models over the Arctic was coordinated with the Arctic Monitoring and Assessment Programme (AMAP, 2015). Comparisons were made for BC and sulphate for six ground stations and during six aircraft campaigns (Eckhardt et al., 2015). As an example, a comparison of the BC concentrations simulated by the ECLIPSE models with measured equivalent BC is shown in Fig. 5 for the stations Zeppelin on Svalbard, Pallas in Finland and Tiksi in Siberia. For Zeppelin, most models clearly underestimate the observed concentrations during winter and spring, whereas for Pallas which is closer to source regions, the models tend to overestimate. In general, the model performance (also at other Arctic sites, not shown) is better than in previous comparisons (e.g., Shindell et al., 2008). However, very large model underestimates were found for Tiksi, from where measurement data have only recently become available. Another ECLIPSE study showed that also the snow BC concentrations are generally underestimated by models in Northern Russia but overestimated elsewhere in the Arctic (Jiao et al., 2014). It is therefore likely that the model underestimates are caused by too low BC emissions in Russia in the ECLIPSE CLE data set. Yttri et al. (2014) attribute this at least partly to an underestimation of residential wood burning, based on levoglucosan measurements made at Zeppelin. Eckhardt et al. (2015) suggest that also SO₂ emissions in Northern Russia are underestimated. ECLIPSE models participating in AMAP (2015) also showed a systematic underestimation in CO concentrations in the Arctic and lack of model skill in simulating reactive nitrogen species important for O₃ production.

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An important finding of the model-measurement comparisons is that overall the ESMs show a similar performance as the CTMs. This is encouraging for the further use of the ESMs for determining the climate impacts (Sect. 3.6). The comparisons led to some further improvements of the ECLIPSE emissions for version 5, prior to their use for transient climate model simulations. For instance, wick lamps were identified as an important emission source in India, the inclusion of which improved the agreement with the observations in a model sensitivity study (Gadhavi et al., 2015). Other enhancements (e.g., re-gridding of non-ferrous smelter emissions to improve SO₂ emissions in Russia as suggested by Eckhardt et al., 2015) came too late for the climate impact studies and were only made in version 5a.

Another aspect of model evaluation is to determine the capability of models to reproduce past trends, and this was tested over Europe. Strong reductions of aerosol emissions occurred over Europe, since the 1980s due to air quality legislation in Western Europe, and since the early 1990s due to economic restructuring in Eastern Europe. This emission reduction is manifest e.g. in strongly increasing trends in surface solar radiation (“solar brightening”) and visibility (Stjern et al., 2011), but also in a stronger warming trend compared to the earlier period in which aerosol emissions increased (Cherian et al., 2014). The “historical” simulations contributed to the 5th Coupled Model Intercomparison Project (CMIP5, Taylor et al., 2012) using previous versions of the ECLIPSE ESMs were assessed for continental Europe, and compared to observations from the Global Energy Balance Archive (Gilgen et al., 1998) and the Climate Research Unit (CRU) of the University of East Anglia (CRU-TS-3.10, Mitchell and Jones, 2005). The 1960–1980 period shows a strong “solar dimming” (reduction in surface solar radiation) and small warming, since the greenhouse-gas-induced warming is offset by the aerosol forcing. The period 1990–2005, in turn, shows the solar brightening, and a much stronger warming. All three tested models are able to reproduce this strong increase in warming trend to within their uncertainties (Fig. 6), suggesting that the climate response to aerosol changes is captured.

3.3 Radiative forcing

To provide input to the metrics needed for designing a mitigation scenario, dedicated model simulations by three ESMs (ECHAM6-HAM2, HadGEM3, NorESM) and a CTM (OsloCTM2) were used to establish a matrix of specific RF (Bellouin et al., 2015) by season and region of emission. Specific RF (SRF) is defined as the RF per unit change in emission rate once the constituents have reached equilibrium and is given in $\text{mW m}^{-2} (\text{Tgyr}^{-1})^{-1}$. To estimate SRF, the emissions of eight short-lived species (BC, OA, SO_2 , NH_3 , NO_x , CO, CH_4 and NMVOCs) were reduced by 20 % compared to their ECLIPSE baseline. These species cause RF themselves and/or lead to the perturbation of radiative forcers (e.g., O_3). The regional reductions were made for Europe and China, as well as for the global shipping sector and for a rest-of-the-world region. To account for seasonal differences in SRF, separate reductions were applied for May–October and November–April. Henceforth we will refer to these as Northern Hemisphere (NH) “summer” and “winter”. Notice that in our case the sign of SRF is opposite to that of RF, because the imposed emission changes are negative. A reduction of a warming species gives negative RF values but positive SRF values. It is important to note that SRF excludes rapid adjustments in the atmosphere, with the exception that BC semi-direct effects were calculated explicitly, and stratospheric temperature adjustments were included for O_3 and CH_4 .

Models generally agreed on the sign of RF and the ranking of the efficiency of the different emitted species, but disagreed quantitatively (see Bellouin et al., 2015 for details). The best estimate of a species’ RF was considered to be the average of all models, with the model spread indicating its uncertainty. However, not all models have calculated RF for all species or have accounted for all processes. For instance, all models were able to quantify the aerosol direct effect but only three quantified the 1st indirect effect. For BC aerosols only one model quantified the snow albedo effect and the semi-direct effect. Therefore, mean RF values were determined by averaging

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across all available models for each process separately. In most cases, all four models were available for this, but for some processes fewer models had to be used.

Figure 7 shows the resulting SRF for reductions in the emissions of SO_2 , NO_x , CH_4 and BC and the processes contributing to the total forcing, for Europe, China, and on global average. The globally-averaged SRF was obtained by adding RF for Europe, China, and Rest of the World, then normalising to global emission change. The SRF values are largest for BC but note that global emissions of BC are smaller than for the other species. In addition, the semi-direct effect of BC potentially offsets a considerable fraction of the aerosol direct RF and RF due to deposition on snow. Quantifying the semi-direct effect has large uncertainties, however, because internal variability of the climate system masks tropospheric adjustments to BC perturbations. This means that the sign of total SRF exerted by decreases in BC emissions may be negative if a weak BC direct effect is more than compensated by a strong semi-direct effect. Nevertheless, the ECLIPSE BC SRF best estimate of about $50 \text{ mW m}^{-2} (\text{Tg}[\text{C}] \text{ yr}^{-1})^{-1}$ when semi-direct effects are included is not an outlier compared to previous estimates, which range from 24 to $108 \text{ mW m}^{-2} (\text{Tg}[\text{C}] \text{ yr}^{-1})^{-1}$ according to Table 23 of Bond et al. (2013). Moreover, ECLIPSE simulations indicate that the magnitude of the semi-direct effect is smaller than the direct aerosol effect (Hodnebrog et al., 2014; Samset and Myhre, 2015), in agreement with most, but not all, previous studies (again, see Table 23 of Bond et al., 2013). Reductions in the emissions of light scattering aerosols such as sulphate stemming from its precursor SO_2 induce a negative SRF. The RF values of aerosols are generally larger for summer emissions than for winter emissions because of the stronger insolation. However, there are exceptions to this. For instance, the BC deposition on snow is more effective for winter emissions because of the larger snow extent in winter and spring and partial preservation of deposited BC into spring. Aerosol SRF is also larger in magnitude for Europe than for China, most likely because of different cloud regimes which are differently affected by semi-direct and indirect aerosol effects.

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Aamaas et al. (2013) investigated several different metrics and showed that emissions of CO_2 are important regardless of what metric and time horizon is used, but that the importance of SLCP varies greatly depending on the metric choices made. MacIntosh et al. (2015) considered the errors made when calculating RF and climate metrics from multi-model ensembles in different ways. They showed that the mean metric values are relatively robust but the estimation of uncertainties is very dependent on the methodology adopted. Finally, Lund et al. (2014a) applied climate metrics to quantify the climate impacts of BC and co-emitted SLCPs from on-road diesel vehicles, and Lund et al. (2014b) considered the special case of a fuel switch from diesel to biodiesel.

As explained in Sect. 3.5, climate metrics were used in ECLIPSE to identify specific sets of air pollution reduction measures that result in net positive climate effects (i.e., reduced warming), considering the impacts on all co-controlled substances. Based on the RF results shown in Sect. 3.3, Aamaas et al. (2015) calculated regional and seasonal GTP and GWP metrics for the SLCP emissions, for various time horizons, and explored their properties. Of all the explored metrics, the pollution control analysis was carried out for GWP_{100} and GTP_{20} , as these two metrics showed large differences in their quantifications. It was found, however, that the emerging basket of emission control measures was very similar for both, although the ranking of the potential climate impacts of individual measures was sometimes different, with larger effect of CH_4 -related measures for the GWP_{100} metric (due to the higher value this metric with longer time horizon assigns to CH_4).

In the following, we concentrate our analysis on the GTP_{20} metric, which is shown in Fig. 8 for a pulse emission relative to an equal mass emission of CO_2 for a selected number of species. The metric builds on and reflects important aspects of the RF forcing values shown in Fig. 7. For example, for most species GTP_{20} values for summer are larger than values for winter (see, e.g., results for SO_2), and values for Europe are larger than values for Asia. While SO_2 only has negative values (i.e., reduction of SO_2 always leads to warming), the opposite is true for CH_4 . Figure 8 also shows the contribution from different processes to the total GTP_{20} metric, and for BC and NO_x they

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measures that reduce the emissions of BC, mainly for small stationary and mobile sources, (3) non-technical measures to eliminate BC emissions, e.g., through economic and technical assistance to the poorest population. While about 220 GAINS model measures were collected in the mitigation basket, they were merged into representative measure groups. For example, high emitters are calculated for each of the GAINS transport sectors and fuels while here removing high emitters is represented by one “measure”. Similarly for cooking stoves, GAINS estimates mitigation potential for various types of fuels but all of these are included further into one category “clean cooking stoves”. Also for CH₄, losses from gas distribution are calculated for several GAINS end use sectors while here the mitigation potential is reported under one measure where leaks from low pressure pipelines are reduced. Finally, for NMVOCs all of the solvent related options in GAINS (over 50) are categorized as one measure reducing losses from solvent use activities. Considering the above, about 50 “measures” represent the about 220 GAINS options that were included in the mitigation basket. The 17 most effective measures contribute 80 % of the total climate benefit, according to the GTP₂₀ metric, with CH₄ measures contributing about 47 % and BC-focused measures contributing 33 %. These measures are listed in Table 3. It is interesting to notice that the top CH₄ and BC-focused measures both concern the oil and gas industry and specifically the venting or flaring of associated gas.

As can be seen in Fig. 2, the mitigation has only minor effects on CO₂ emissions, but reduces most SLCPs strongly compared to the CLE scenario. By 2030, CH₄ emissions are reduced by about 50 % and BC emissions by nearly 80 %. OA is co-controlled with BC, causing a nearly 70 % (not shown) reduction of its emissions, as for some sectors BC outweighs the cooling effects of OA. While NO_x emission reductions are in most cases also not preferred by the GTP₂₀ metric (see Fig. 8), reductions stem from the co-control when higher Euro standards are introduced; they reduce significantly several pollutants such as BC, CO, NMVOC and also NO_x (Klimont et al., 2015a, b). By contrast, SO₂ emissions are nearly the same in both the CLE and MIT scenarios, as for the key sectors emitting SO₂, the warming by SO₂ reductions (see Fig. 8) cannot

be outweighed by co-control of species whose reduction would lead to cooling. Thus, SO₂ reductions are largely avoided.

The global CO₂-equivalent emissions (calculated using the GTP₂₀ metric values) are shown as a function of time in Fig. 9. Values are shown for CLE and MIT and are split into contributions from CH₄ and other SLCP emissions. For comparison, CO₂ emissions are also shown. On the short time scale of the GTP₂₀ metric, CH₄ and CO₂ emissions are nearly equally important in the CLE scenario. The CO₂-equivalent emissions of CH₄ are, however, reduced by 50 % in MIT. The CO₂-equivalent emissions for the other SLCPs are negative in both scenarios, indicating that in total they have a cooling impact. As the mitigation reduces preferentially warming components, this cooling becomes even stronger in the MIT case. The total CO₂-equivalent emissions (including CO₂ emissions) are reduced substantially in the MIT scenario (blue shaded area in Fig. 9 shows the reduction), for example by about 70 % in the year 2030. About 56 % of this reduction is due to CH₄ measures and ~ 44 % is due to other measures. It is important to notice that the effect of the SLCP mitigation is relatively large for the rather short time horizon of the GTP₂₀ metric; it would be smaller for a longer time horizon. Similarly, the relative importance of CH₄ compared to other SLCP emissions would be increased for a longer time horizon.

Surface concentrations of SLCPs resulting from the CLE and MIT scenarios were determined from the various model simulations. Figure 10 shows maps of the model-mean relative differences for O₃ and PM_{2.5} for the last decade (2041–2050) of the transient simulations and Table 4 reports differences for several regions shown with boxes in Fig. 10. Concentrations of O₃ (Fig. 10, upper panel) are reduced globally, with reductions of more than 12 % in most of the Northern Hemisphere and the strongest reductions of up to about 20 % occurring in East Asia. For instance, in Eastern China (see Table 4), O₃ is reduced by 19.3 % (16.0–24.4 %). BC and OA concentrations (not shown) were also globally reduced, with BC reductions reaching more than 80 %. For sulphate (not shown), the relative changes are much smaller than for BC and OA and both increases and decreases occur – a consequence of the relatively small global

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SO₂ emission reductions (see Fig. 2). Changes in PM_{2.5} concentrations (Fig. 10, lower panel) are smaller because of large contributions from natural sources (e.g., sea salt, dust, wildfire emissions). PM_{2.5} concentrations in the SLCP source regions were reduced by typically 10–20% and up to nearly 50% in smaller regions. Reductions are strongest in Asia, for instance 19.8% (17.9–22.5%) in India (Table 4).

In summary, air pollutant concentrations in the MIT scenario are dramatically reduced compared to the CLE scenario, especially in the polluted (and heavily populated) source regions. This indicates the beneficial effect of the SLCP mitigation on air quality. Nevertheless, a word of caution is needed for O₃. The O₃ concentrations increase strongly (typically between 5 and 20%, depending on region and season) between now and 2050 in the CLE scenario, because of increasing CH₄ and NO_x emissions. Therefore, global mean O₃ concentrations even in the MIT scenario do not decrease substantially with time. However, strong relative O₃ reductions by the mitigation are simulated in the SLCP source regions (Table 4), which more than outweigh the overall concentration increase in the CLE scenario. Therefore, population exposure to O₃ decreases with time in the MIT scenario.

For Europe and Asia, the GAINS model also contains source-receptor relationships which allow the estimation of the impacts of emissions on human health. GAINS quantifies the impacts of changes in SLCP emissions on the long-term population exposure to PM_{2.5} in Europe, China and India and estimates the resulting premature mortality, in terms of reduced statistical life expectancy and cases of premature deaths (Amann et al., 2011). Calculations follow the recommendations of the findings of the WHO review on health impacts of air pollution and recent analyses conducted for the Global Burden of Disease project (Lim et al., 2012), relying on the results of the American Cancer Society cohort study (Pope et al., 2002) and its re-analysis (Pope et al., 2009). It uses cohort- and country-specific mortality data extracted from life table statistics to calculate for each cohort the baseline survival function over time.

Using GAINS, we estimate that in the EU the loss of statistical life expectancy will be reduced from 7.5 months in 2010 to 5.2 months in 2030 in the CLE scenario. The

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ECLIPSE mitigation measures (MIT) would reduce statistical life shortening by another 0.9 months (Fig. 11, upper panel), resulting in 4.3 months of reduction in life expectancy. This value is only slightly above the target of 4.1 months that has been set by the European Commission in its 2013 Clean Air Policy proposal (EC 2013). Population in non-EU countries would gain approximately one month life expectancy from the implementation of the ECLIPSE measures in 2030 (Fig. 11, upper panel).

In China and India, the potential health gains from the implementation of the ECLIPSE measures are significantly larger, however, starting from a substantially higher level of life shortening due to $PM_{2.5}$ (Fig. 11, lower panel). In China, the ECLIPSE measures would in the year 2030 extend the life expectancy of the population by approximately 1.8 months and reduce the premature deaths attributable to $PM_{2.5}$ by 150 000–200 000 cases year⁻¹.

In India, rapid increase in energy consumption, together with lacking regulations on emission controls for important sources (e.g., power generation) and poor enforcement of existing laws (e.g., for vehicle pollution controls) will lead to a steep increase in $PM_{2.5}$ levels. If no saturation of health impacts is assumed for such high levels (there are no cohort studies available for such high concentrations), with conservative assumptions GAINS estimates approximately 850 000 cases annually of premature deaths from air pollution in 2010. For 2030, $PM_{2.5}$ exposure would increase by more than 50 %, and at the same time the population would increase and age. Combined, these factors would let the premature deaths from air pollution grow by approximately 125 % to 1.9 million cases in 2030, with another doubling to 3.7 million cases in 2050. Against this background, the ECLIPSE measures would avoid more than 400 000 cases of premature deaths in 2030 and almost 700 000 in 2050. Using the loss in statistical life expectancy as an alternative metric, the ECLIPSE measures would gain 11–12 months to the Indian population (Fig. 11, lower panel).

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mean. A large part of this increase is a response to increasing CO₂ concentrations, and consequently the MIT scenario also showed an (albeit smaller) temperature increase. As we here are mainly interested in the response to the SLCP mitigation, we only consider the difference between the MIT and the CLE scenario in the following. Time series of the global mean temperature difference between the two scenarios are shown in Fig. 14. There is a considerable spread between individual ensemble members even from the same model. This reflects simulated natural climate variability superimposed on the response to the SLCP mitigation, which makes diagnosis of the latter difficult. Systematically negative global mean temperature responses emerge only when the mitigation measures are fully implemented. The multi-model mean global temperature (MIT-CLE) difference is -0.22 ± 0.07 K for the final 10 years, confirming that the mitigation could be successful in reducing the warming of the CLE scenario.

The relative cooling is particularly strong over the continents and weakest over the Northern North Atlantic (not shown), similar to the response patterns seen in the perturbation simulations. Figure 15 shows mean temperature responses for the last decade of the simulation, for various regions. The strongest relative cooling between MIT and CLE of about 0.44 K (0.39–0.49 K) are found for the Arctic, with peak values of about 0.62 K (0.37–0.84 K) occurring in autumn (winter values are similar). Over Europe, differences are more consistent between models and more strongly negative in the southern parts than in the northern parts (see also Fig. 16). This can be explained by the larger natural climate variability in Northern Europe. Mainly small and inconsistent results are found over India, due to model differences in the shift of the ITCZ, whereas changes over China and North America are consistently negative from (almost) all models and for all seasons.

The precipitation responses are less robust which, in the tropics, is due to model differences in the migration of the ITCZ. Nevertheless, there are regions with consistent responses of all models. Of particular interest is the precipitation increase over Southern Europe (Fig. 16). Seen against the background of expected warming and drying in the Mediterranean area due to CO₂-driven climate change, the precipitation

contributions from individual SLCP species. Notice that it would be prohibitively expensive to run an ensemble of ESM simulations that is large enough to detect the small temperature response resulting from such minute emission changes.

4 Conclusions

ECLIPSE has come to a number of important scientific conclusions, which are also of high relevance for climate and air quality policy:

- ECLIPSE has created a new inventory for anthropogenic SLCP emissions, including scenarios for the future. An important finding is the large range of possible future developments of anthropogenic SLCP emissions, which even for a single future energy pathway substantially exceeds the range of SLCP emissions given in IPCC's RCPs. The large range results from the uncertainties of future air quality policies, as well as from the expected level of implementation and enforcement of existing policies (Klimont et al., 2015a, b).
- Detailed comparisons between measured and modelled distributions of aerosol, O₃ and other SLCP gases have shown that for many substances the models are in good agreement with available background observations. The model performance of the ESMs is similar to that of CTMs. For BC, in particular, the agreement between models and measurements has improved for the Arctic (Eckhardt et al., 2015), which is partly the result of accounting for emissions from gas flaring and emission seasonality (Stohl et al., 2013). Outside the Arctic, a reduction of the BC lifetime led to improvements (Samset et al., 2014). Nevertheless, our comparisons suggest underestimates of BC and aerosol precursor emissions in high-latitude Russia and in India. Furthermore, it was found that SO₂ concentrations are overestimated and CO concentrations are underestimated by the models (Quennehen et al., 2015) at the surface in Asia and Europe during summer and autumn. The CO underestimate is likely associated with a too short CO lifetime in the mod-

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els. Ozone, on the other hand, is generally overestimated at rural locations. Such discrepancies may affect model responses to emission perturbations and thus radiative forcing.

- Earth System Models can reproduce the accelerated upward trend of surface temperature over Europe that was observed when aerosol precursor emissions were reduced in the 1990s (leading to solar brightening), after a period of emission growth in the 1960s–1980s (leading to solar dimming) (Cherian et al., 2014).
- ECLIPSE performed detailed multi-model calculations of RF for all considered SLCP species, as a function of emission region and season. It is found that the absolute values of specific RF for aerosols are generally larger in summer than in winter (Bellouin et al., 2015). It is found that the semi-direct effect on clouds, although highly uncertain, can potentially offset a considerable fraction of the direct positive RF of BC. This, together with reduced BC lifetimes, causes the net RF for BC calculated in ECLIPSE to be only weakly positive, which is different from most previous studies.
- NO_x emissions affect the concentrations of O_3 , CH_4 and nitrate aerosols. The first effect leads to positive RF, while the latter two cause negative RF. We have quantified all these effects and can state with confidence that the current net RF of global historical NO_x emissions is negative. The forward looking metrics GWP and GTP for NO_x are negative as well, except for short time horizons.
- ECLIPSE had a focus on calculation and testing of emission metrics, which led to a better understanding of existing metrics (Aamaas et al., 2013), further development of the applications of the RTP concept (Collins et al., 2013) and introduction of new metric concepts such as the Global Precipitation change Potential (GPP) by Shine et al. (2015). After careful consideration of the alternatives, we chose a 15 year ramp-up version (i.e., assuming a linear implementation of measures) of the GTP_{20} metric for designing a SLCP mitigation scenario.

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ing for OA removal (Baker et al., 2015a). The global mean temperature response to BC removal was slightly negative: -0.05 K (-0.15 to 0.08 K). The relatively small global response to BC emission reductions was attributed to strong, while uncertain, indirect and semi-direct effects, which partly offset the direct aerosol radiative effect.

- Climate impacts of the MIT scenario were investigated with ESM ensemble transient simulations of both the CLE and MIT scenario (Baker et al., 2015b). Multi-model ensemble mean global mean surface temperature in the CLE scenario increased by 0.70 ± 0.14 K between the years 2006 and 2050. The ensemble mean global mean surface warming for the last decade of the simulation (2041–2050) was, however, 0.22 ± 0.07 K weaker for the MIT scenario, demonstrating the effect of the SLCP mitigation. The response was strongest in the Arctic, with warming reduced by about 0.44 K (0.39 – 0.49 K).
- In addition to global annual mean temperature change, there are other climate parameters that are of relevance for policy decisions (e.g., changes in precipitation, regional temperatures, etc.). The SCLP reductions in the MIT scenario led to particularly beneficial climate responses in Southern Europe, where the surface warming was reduced by about 0.3 K from spring to autumn and precipitation rates were increased by about 15 (6 – 21) mm yr^{-1} (15 mm yr^{-1} corresponding to more than 4 % of total precipitation), compared to the CLE scenario. Thus, the mitigation could help to alleviate expected future drought and water shortages in the Mediterranean region.
- Additional ESM transient simulations, which only included the CH_4 emission reductions, led to a global warming reduction that amounted to about 90 % of the reduction produced by the simulations using the full set of measures for the final decade of the simulations (2041–2050). This suggests that, for longer time horizons, the net climate benefits from our chosen non- CH_4 SLCP mitigation measures in terms of global annual mean temperature change are limited, probably

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0.01–0.06 K for time periods from 2021–2040, according to metrics-based estimates). Thus, to maximize climate co-benefits of non-CH₄ SLCPs, sources with the highest BC/OA emission ratios should be addressed with priority. At the same time, air pollution policies should consider mitigation of CH₄, with clear co-benefits for climate warming and air quality via reduced surface O₃ concentrations.

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Table 1. Air quality standards for Europe (European Union reference values), WHO air quality guidelines (AQG), US-EPA National Ambient Air Quality Standards (NAAQS) and the Environmental Quality Standards (EQS) and guideline values for air pollutants in Japan. Values in brackets give time period for which the guideline is defined.

Pollutants	EU reference levels ^a	WHO AQG ^b	USEPA NAAQS ^c	Japan EQS ^d
PM _{2.5}	20 µg m ⁻³ (year)	10 µg m ⁻³ (year)	12 µg m ⁻³ (year)	15 µg m ⁻³ (year)
PM ₁₀	4 µg m ⁻³ (day)	20 µg m ⁻³ (year)	150 µg m ⁻³ (day)	100 µg m ⁻³ (day, SPM ^e)
O ₃	120 µg m ⁻³ (8 h)	100 µg m ⁻³ (8 h)	0.075 ppm (8 h)	118 µg m ⁻³ (1 h ^f)
NO ₂	40 µg m ⁻³ (year)	40 µg m ⁻³ (year)	53 ppb (year)	75–113 µg m ⁻³ (1 h)
SO ₂	125 µg m ⁻³ (day)	20 µg m ⁻³ (day)	75 ppb (1 h)	105 µg m ⁻³ (1 day)
CO	10 mg m ⁻³ (8 h)	10 mg m ⁻³ (8 h)	9 ppm (8 h)	10 ppm (1 h)

^a EEA 2013, Indicator CSI 004.

^b WHO Air Quality Guidelines (WHO, 2006).

^c US-EPA National Ambient Air Quality Standards (<http://www.epa.gov/air/criteria.html#3>, last access: 16 April 2014).

^d Environmental Quality Standards (EQS) and guideline values for air pollutants in Japan (Kawamoto et al., 2011).

^e 100 % efficiency cut-off at 10 µm while PM₁₀ is defined as 50 % efficiency cut-off at 10 µm aerodynamic diameter (Kawamoto et al., 2011).

^f Photochemical oxidants (O_x) (Kawamoto et al., 2011).

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Table 2. Overview of the ECLIPSE models and how they were set up for the years 2008–2009.

Model Name	Model Type*	Horizontal/vertical resolution	Meteorological fields	Periods simulated/output temporal resolution	References
FLEXPART	LPDM	Meteorological input $1^\circ \times 1^\circ$, 92L	ECMWF Operational Analyses	2008–2009, 3 h	Stohl et al. (1998, 2005)
OsloCTM2	CTM	$2.8^\circ \times 2.8^\circ$, 60L	ECMWF Operational Analyses	2008–2009, 3 h	Myhre et al. (2009), Skeie et al. (2011)
EMEP	CTM	$1^\circ \times 1^\circ$, 20L	ECMWF operational	2008–2009, 24 h	Simpson et al. (2012)
TM4-ECPL	CTM	$2^\circ \times 3^\circ$, 34L	ECMWF ERA-interim	2008–2009, 24 h	Kanakidou et al. (2012), Daskalakis et al. (2015)
WRF-CMAQ	CTM	50 km \times 50 km, 23L	NCEP	2008, 24 h	Im et al. (2013)
WRF-Chem	CTM	50 km \times 50 km, 49L	Nudged to FNL	Mar–Aug 2008, 3 h	Grell et al. (2005), Zaveri et al. (2008)
NorESM	ESM	$1.9^\circ \times 2.5^\circ$, 26L	Internal, observed SST prescribed	2008–2009, 3 h	Kirkevåg et al. (2013), Bentsen et al. (2013)
ECHAM6-HAM2	ESM	$1.8^\circ \times 1.8^\circ$, 31L	ECMWF Reanalysis	Mar–Aug 2008, 1 h	Stevens et al. (2013), Zhang et al. (2012)
HadGEM3	ESM	$1.9^\circ \times 1.3^\circ$, 63L	ECMWF ERA-interim	Mar–Jun, Nov 2008, Jan, May and Nov 2009, 2 h	Hewitt et al. (2011), Mann et al. (2010)
CESM-CAM4	ESM	$1.9^\circ \times 2.5^\circ$, 26L	Internal	Was not evaluated for 2008–2009; only used for 2000–2050 simulations	Gent et al. (2011)

* Chemistry transport model (CTM), Lagrangian particle dispersion model (LPDM), Earth system model (ESM).

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Table 3. Top-17 mitigation measures contributing together more than 80 % of the climate benefit. Measures are ranked by importance starting from the top. Methane measures contribute about 47 % of the benefits according to the GTP₂₀ metric, while 33 % are attributed to BC focused measures; 20 % are contributed by measures not listed here.

Methane measures	Measures targeting BC reduction
Oil and gas industry: Recovery and use (rather than venting or flaring) of associated gas	Oil and gas industry: Improving efficiency and reducing gas flaring
Oil and gas industry (unconventional): Reducing emissions from unintended leaks during production (extraction) of shale gas	Transport: Eliminating high emitting vehicles (super-emitters)
Coal mining: Reducing (oxidizing) emissions released during hard coal mining (ventilation air CH ₄)	Residential-commercial: Clean biomass cooking stoves
Waste: Municipal waste – waste paper separation, collection, and recycling	Residential-commercial: Replacement of kerosene wick lamps with LED lamps
Waste: Municipal food waste separation, collection and treatment in anaerobic digestion (biogasification) plants	Transport: Widespread Euro VI emission standards (incl. particle filters) on diesel vehicles
Coal mining: Hard and brown coal -pre-mining emissions – Degasification	Industrial processes: Modernized (mechanized) coke ovens
Gas distribution: Replacement of grey cast iron gas distribution network	Agriculture: Effective ban of open field burning of agricultural residues
Waste: Industrial solid waste (food, wood, pulp and paper, textile) – recovery and incineration	
Waste: Wastewater treatment from paper and pulp, chemical, and food industries – anaerobic treatment in digester, reactor or deep lagoon with gas recovery, upgrading and use. For residential wastewater centralized collection with anaerobic secondary and/or tertiary treatment (incl. treatment with bacteria and/or flaring of residual CH ₄)	
Oil and gas industry (conventional): Reducing emissions from unintended leaks during production (extraction)	

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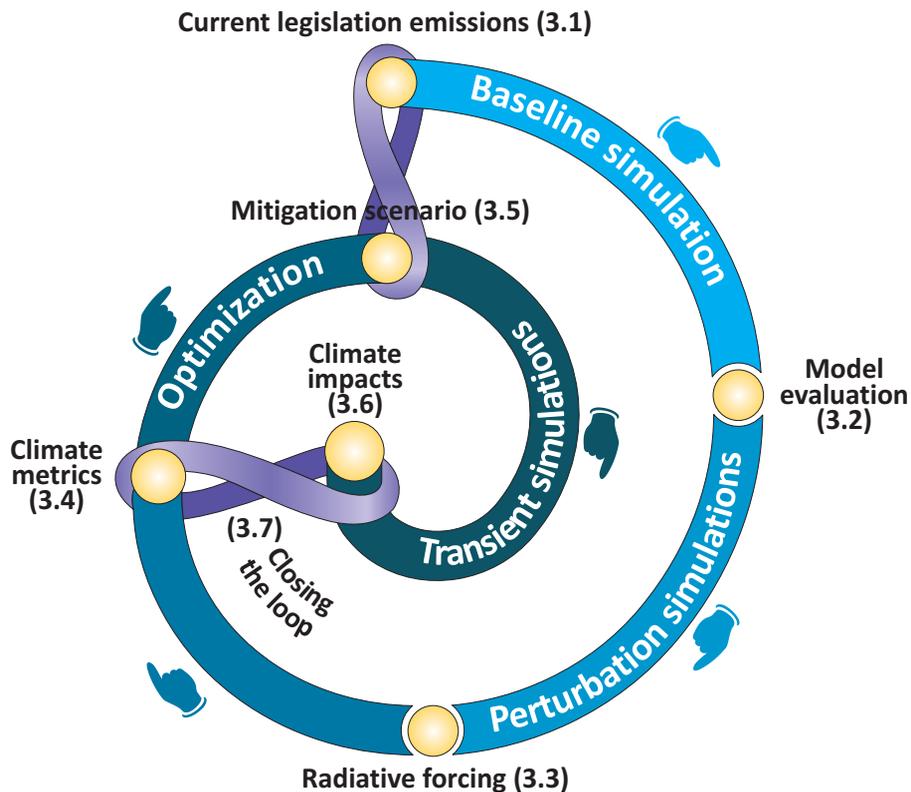


Figure 1. Schematic of the ECLIPSE overall methodology. Numbers in brackets correspond to section numbers in this paper.

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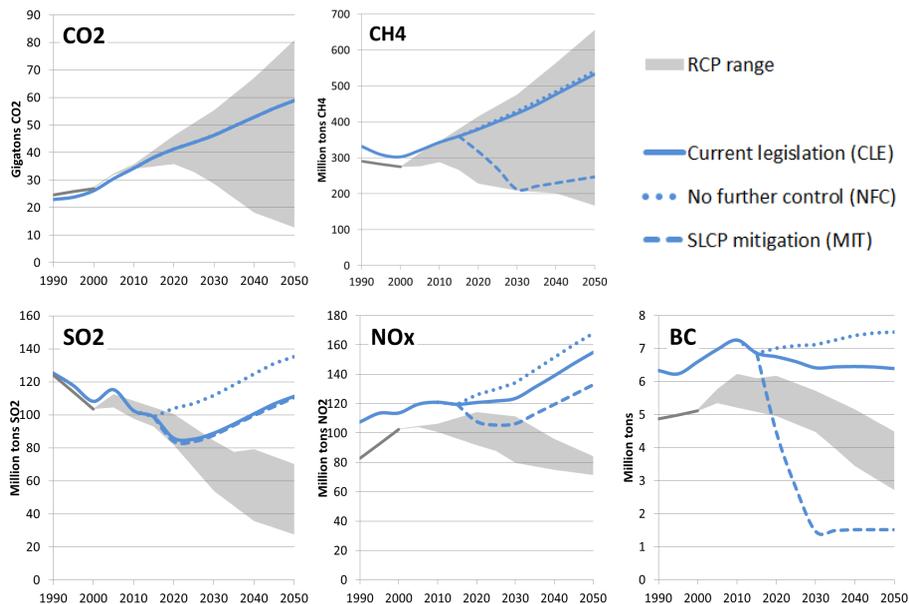


Figure 2. Global annual anthropogenic emissions of CO₂, CH₄ and key air pollutants (SO₂, NO_x and BC) for the current legislation (CLE), no further controls (NFC) and ECLIPSE SLCP mitigation scenario. Units are Gt for CO₂ and Mt for the SLCPs. Also shown for comparison is the range of the RCP emission scenarios (grey shading).

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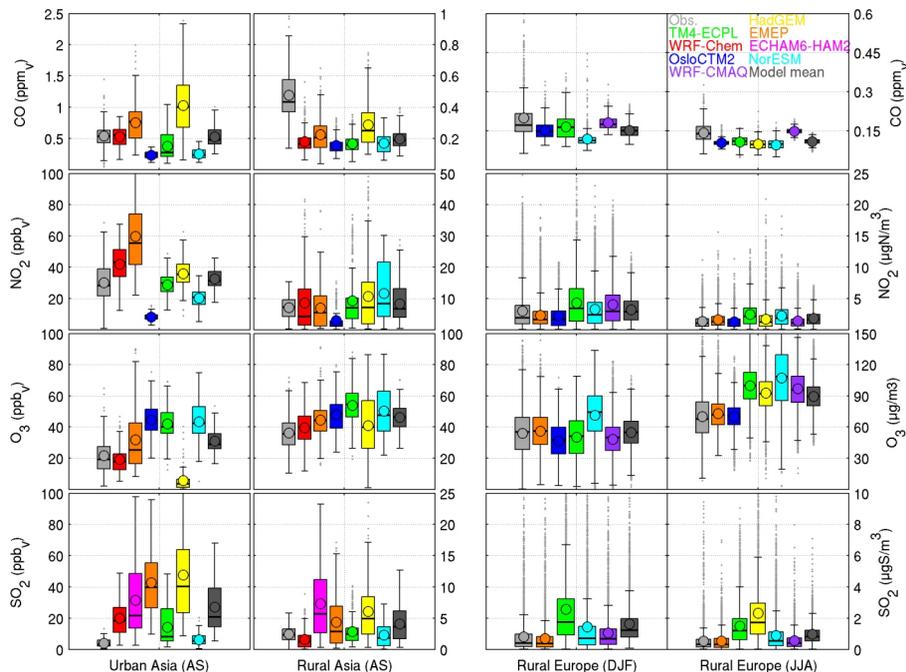


Figure 3. Box and whiskers plots showing the frequency distribution of measured and modelled CO, NO₂, O₃ and SO₂ mixing ratios or concentrations representative for background stations in urban and rural areas in East Asia during August and September 2008 (two left panel columns) and for rural background stations in Europe for winter (December–February, DJF) and summer (June–August, JJA) 2008 (two right panel columns). Circles and central lines show the means and the medians, respectively; box edges represent the 25th and the 75th percentiles. For East Asia, results are averaged over several sites: Beijing, Inchon and Seoul for the urban areas, Gosan, Kunsan, Kangwha, Mokpo and Taean for rural areas. Results for individual sites can be found in Quennehen et al. (2015). For Europe, daily mean observed values are averaged over all stations of the EMEP network with available data. Model data are treated like the observations and only the days with available observations are taken into account.

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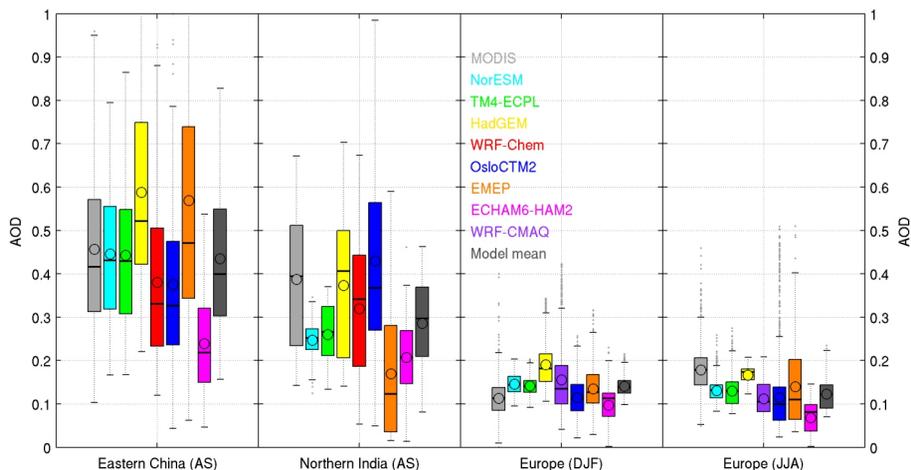


Figure 4. Comparison of satellite-derived (MODIS) and modelled aerosol optical depth (AOD) at wavelengths of 550 nm over Eastern China and Northern India (for area definition, see Quenhen et al., 2015) in August–September 2008, and Europe (14.5° W–34.5° E, 35.5–74.5° N) in winter (December–February, DJF) and summer (June–August, JJA) of 2008. Mean values (circles), medians (central lines), 25th and 75th percentiles (boxes) and range of other data excluding outliers (whiskers) are shown.

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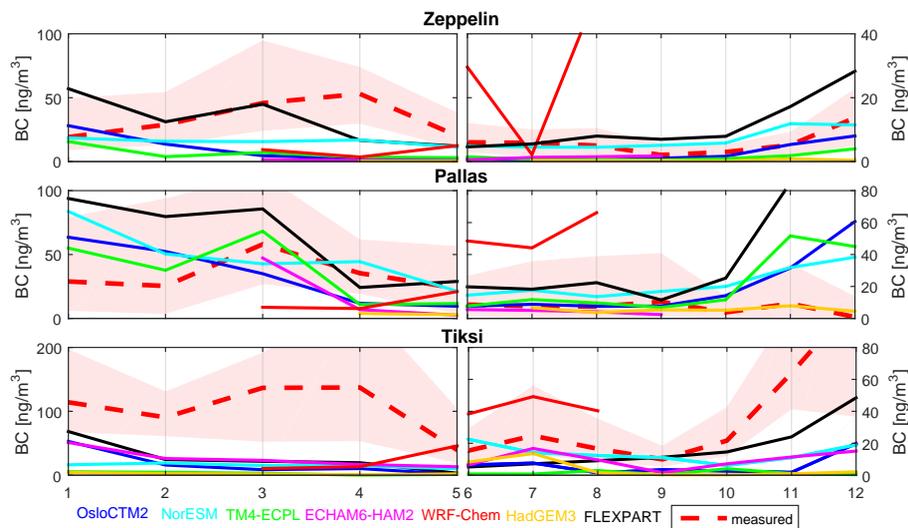


Figure 5. Monthly (month is displayed on the abscissa) median observed and modelled BC concentrations for the stations Zeppelin on Svalbard (11.9° E, 78.9° N, top), Pallas in Finland (24.12° E, 67.97° N, middle) and Tiksi in Siberia (128.9° E, 71.6° N, bottom), for late winter/spring (left column) and summer/autumn (right column) for the years 2008–2009 (for Tiksi, measured values were available only from July 2009 to June 2010). The red dashed lines connect the observed median values, the light red shaded areas span from the 25th to the 75th percentile of the observations. Modelled median values are shown with lines of different colour according to the legend. Notice that different concentration scales are used for individual panels and also for January–May (axis on left hand side) and June–December (axis on right hand side) periods. Modified from Eckhardt et al. (2015).

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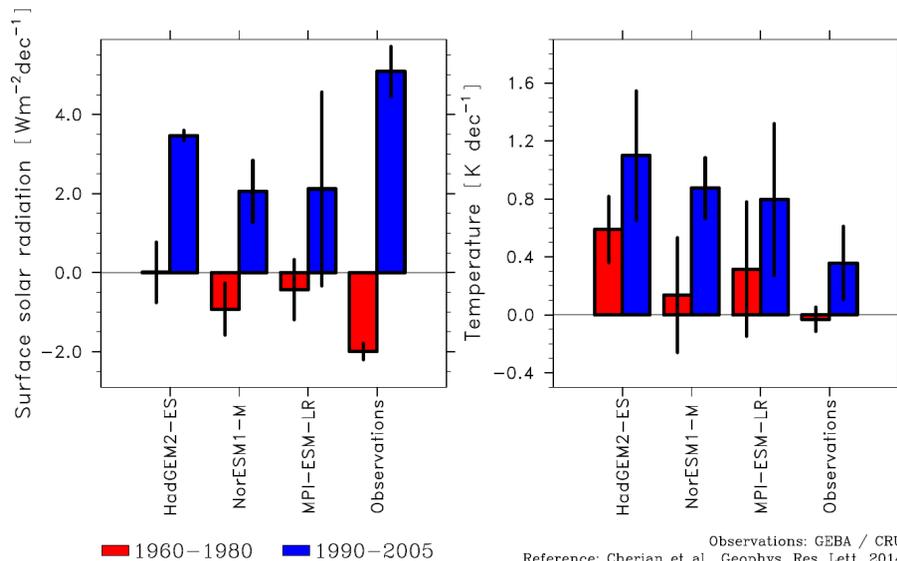


Figure 6. Linear trends in (left) surface solar radiation and (right) near-surface temperature increase per decade over continental Europe from the “historical” simulations in the CMIP5 archive contributed by previous versions of three ECLIPSE ESMs. The period 1960–1980 is shown in red, the period 1990–2005 in blue.

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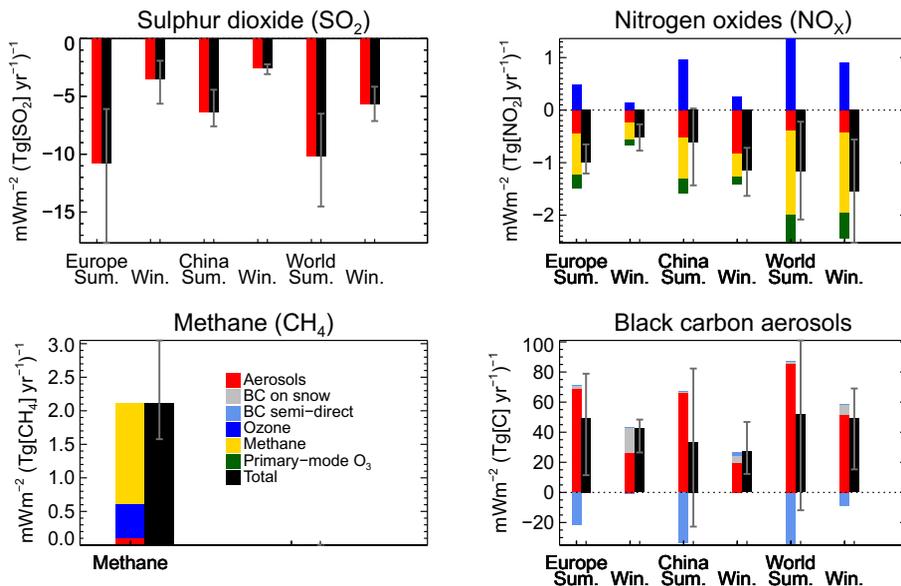


Figure 7. The ECLIPSE estimates of specific radiative forcing (SRF, $\text{mW m}^{-2} \text{Tg}^{-1} \text{yr}^{-1}$ of emission rate change) for reductions in the emissions of SO₂, NO_x, CH₄, and BC aerosols in Europe, China and the global average, separately for NH summer (Sum., May–October) and NH winter (Win., November–April). Shown are values averaged over all five models, with the error bars indicating the full range of the model estimates. Colours indicate the contribution of different forcing mechanisms. Notice that for CH₄ regionality was not accounted for because, due to its longer lifetime, forcing does not depend on the emission region.

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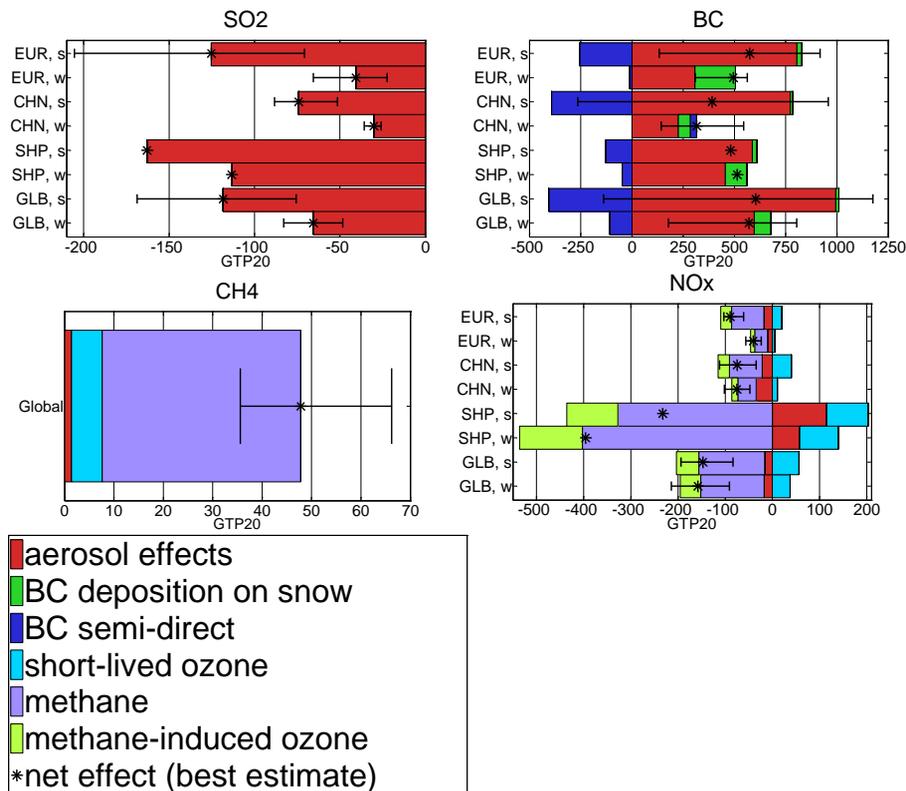


Figure 8. GTP_{20} values for SLCPs, relative to an equal mass emission of CO_2 , for all regions and seasons, decomposed by processes and based on the RF values shown in Fig. 7. The regions included are Europe (EUR), China (CHN), global (GLB), and the shipping sector considered separately (SHP), all for both NH summer (s, May–October) and NH winter (w, November–April). Uncertainty bars reflect model spread. Only two models calculated effects of emissions from shipping and there no uncertainty range is given.

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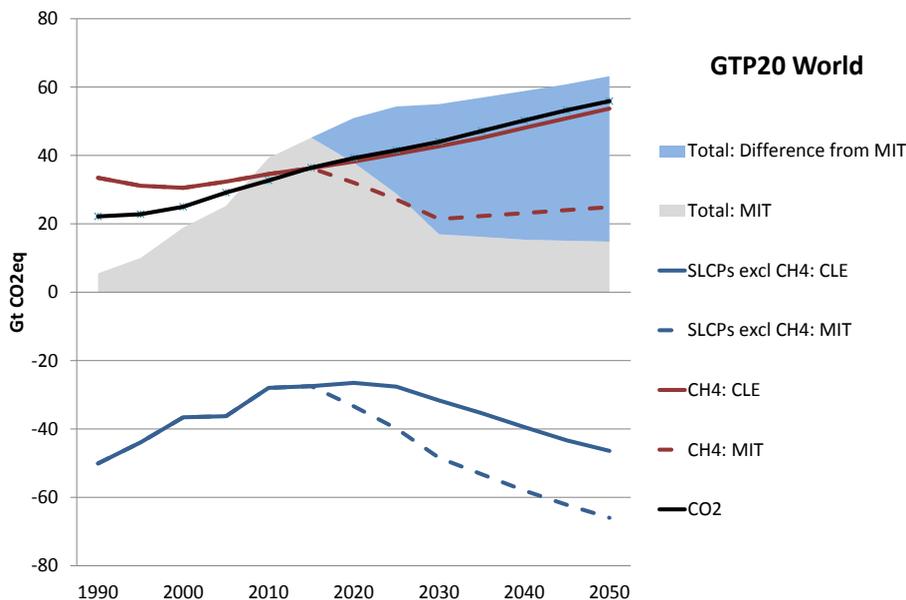


Figure 9. CO₂-equivalent emissions (Gt) determined with the GTP₂₀ metric, as a function of time, for global emissions of the current legislation (CLE) and SLCP mitigation (MIT) scenarios. Lines show the values for individual forcing components (black CO₂, red CH₄, blue other SLCPs), while the shaded areas show the total CO₂ equivalent emissions from all SLCPs and CO₂. The blue shading indicates the mitigation potential of MIT, compared to CLE.

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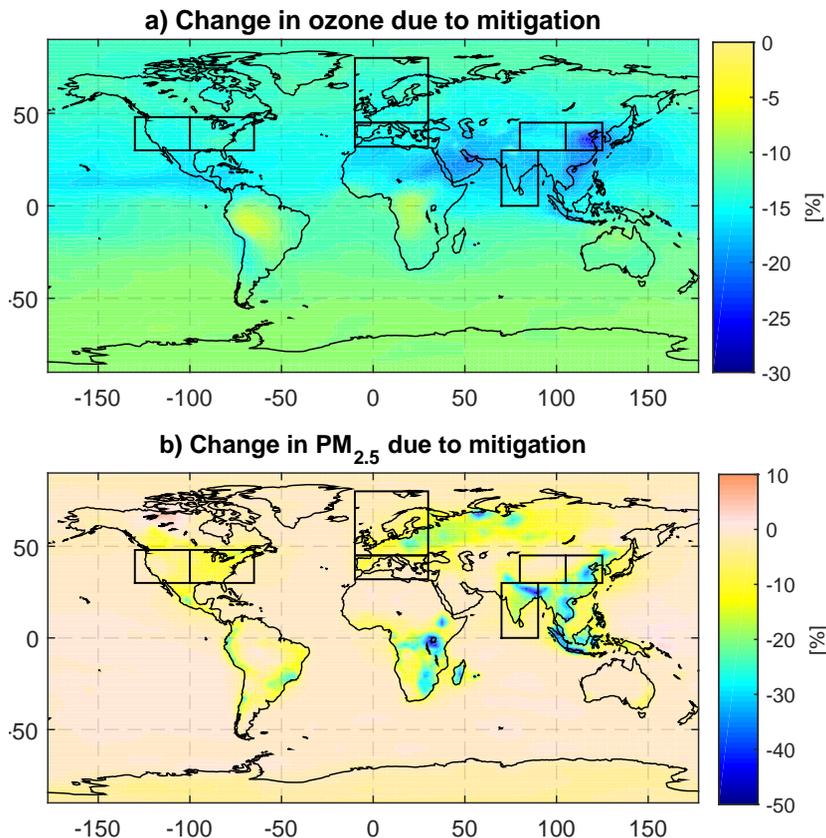


Figure 10. Relative difference maps (in %) for the surface concentrations of O₃ and PM_{2.5}, as obtained from simulations based on the emission mitigation (MIT) scenario and the current legislation (CLE) scenario, i.e. $(100 \times (\text{MIT}-\text{CLE})/\text{CLE})$. Shown are mean concentration differences for the period 2041–2050 and averaged over model results from OsloCTM2, NorESM and HadGEM. The black boxes define the regions used in Table 4.

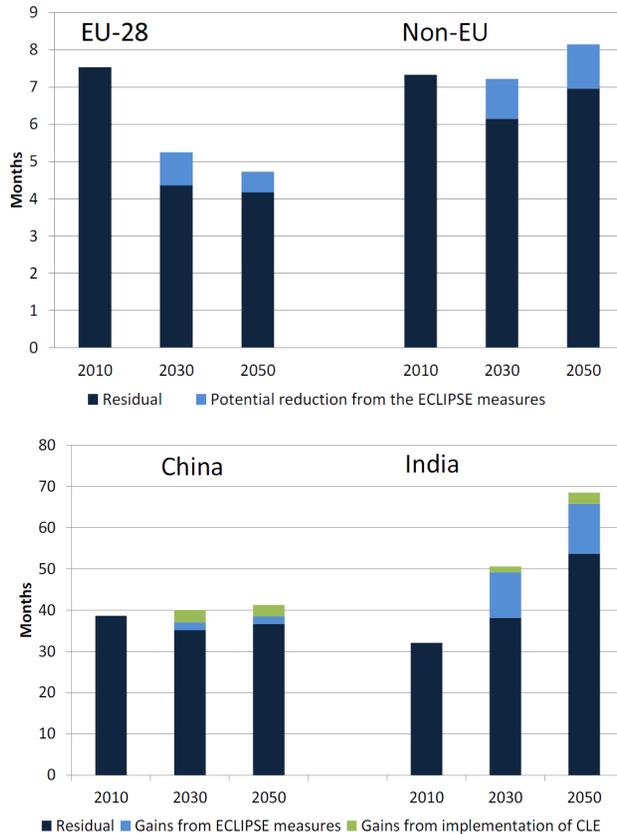


Figure 11. Loss of statistical life expectancy (months) due to the exposure to $PM_{2.5}$ in Europe, for EU-28 and Non-EU countries (upper panel) and for China and India (lower panel). The black bars give the values for the mitigation (MIT) scenario, whereas the blue increments show the difference to the current legislation (CLE) scenario. For India and China, green bars indicate the gains from the implementation of the CLE scenario.

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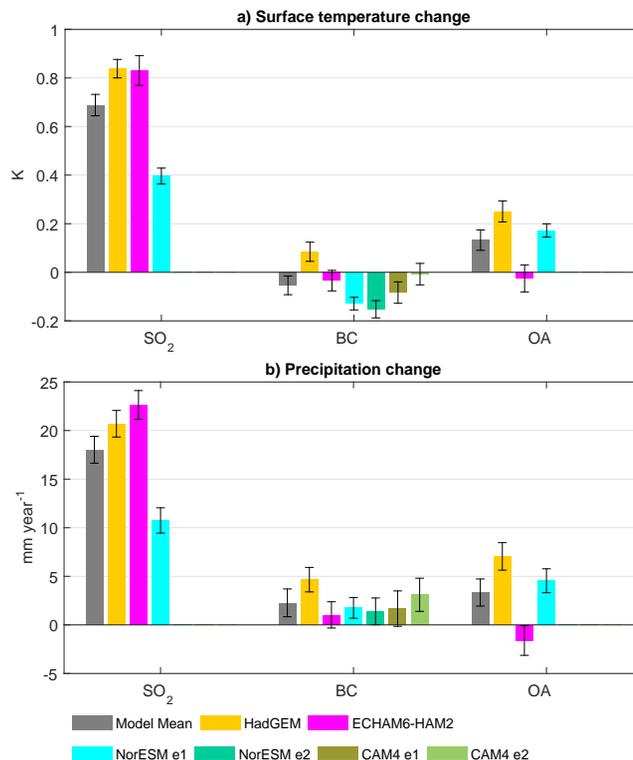


Figure 12. Global mean annual average changes in **(a)** surface temperature and **(b)** precipitation averaged over all land areas excluding ice sheets, from the four ESMS (one was run only for BC), for a complete removal of all land-based emissions of a particular species, compared to the ECLIPSE version 4 baseline emissions for the year 2008 (CESM-CAM4 used the year 2000). The plot shows results averaged over 50 year model simulations; for NorESM and CESM-CAM4, two ensemble members each were run for BC. Error bars are 95 % confidence intervals in the mean, based on 50 annual means from each model and thus only reflect the uncertainty of the mean caused by natural variability within each model.

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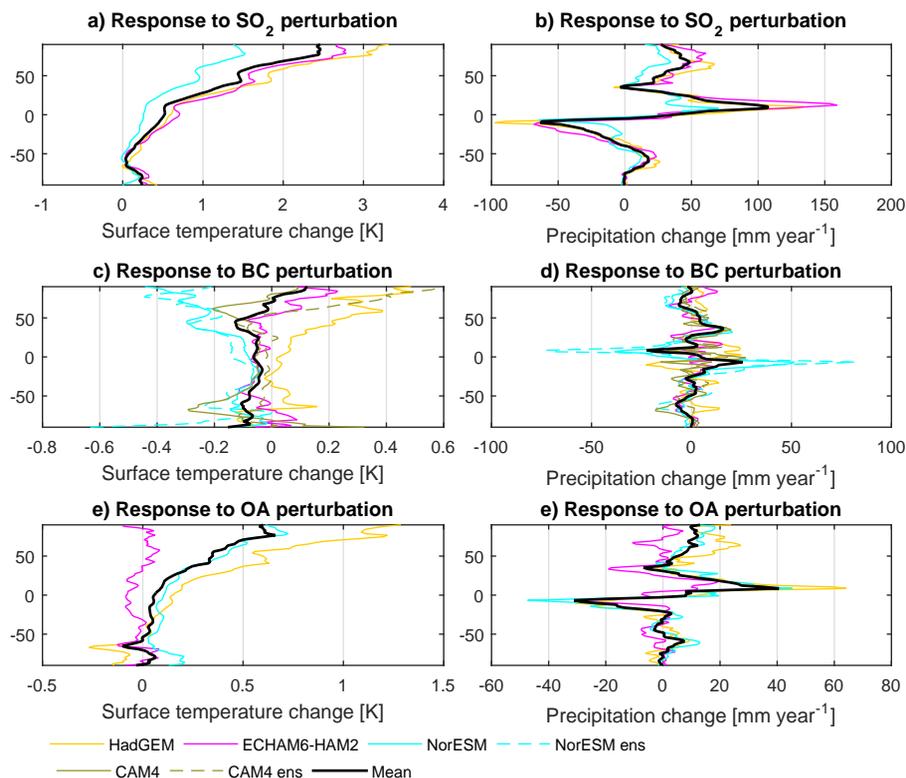


Figure 13. Annual mean changes in zonal-mean surface temperature (left panels **a**, **c**, **e**) and precipitation (right panels **b**, **d**, **f**) for a complete removal of all land-based emissions of (**a**, **b**) SO₂, (**c**, **d**) BC, and (**e**, **f**) OA. Notice the differences in scales between different panels.

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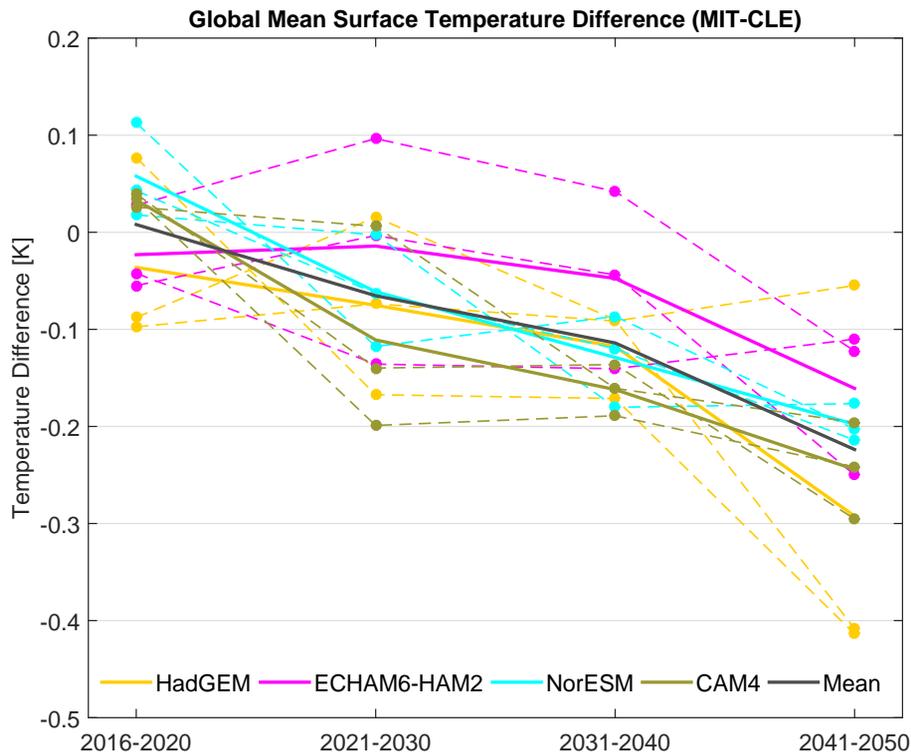


Figure 14. Time evolution of differences in global mean surface temperature between transient simulations following the mitigation (MIT) and the current legislation (CLE) scenario, i.e. (MIT-CLE), for the four ECLIPSE models. Negative values mean that temperatures are lower in the MIT than in the CLE scenario. The ensemble means for each model are shown as thick lines, whereas the individual ensemble members are shown with thin lines.

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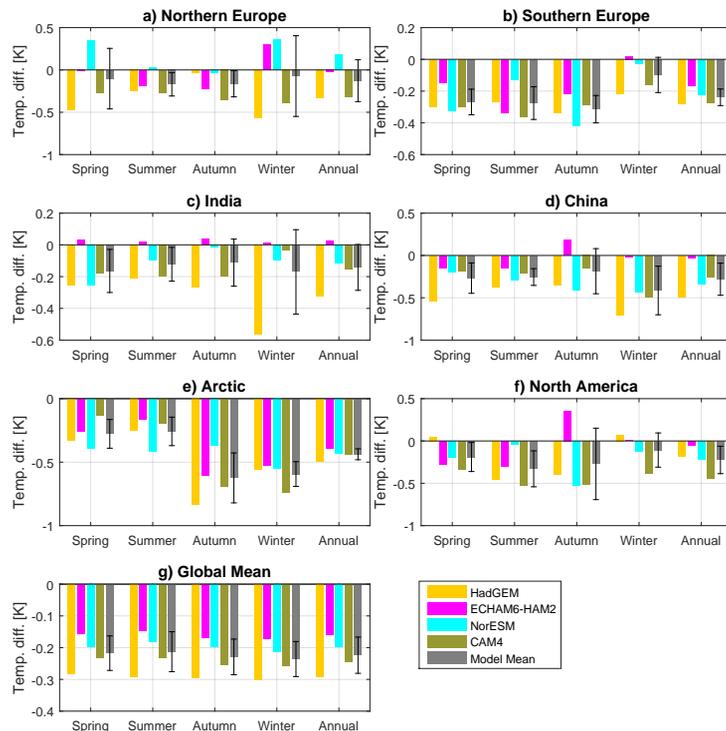


Figure 15. Seasonal and annual mean differences in surface temperatures (in K) in various regions (a–f) and for the whole globe (g) between transient simulations of the mitigation (MIT) and the current legislation (CLE) scenario, i.e. (MIT-CLE), averaged over the last 10 years of the simulations (2041–2050). Regions are defined as (a) 45–65° N, 10° W–65° E, (b) 30–45° N, 10° W–65° E, (c) 7–35° N, 68–90° E, (d) 24–48° N, 80–132° E, (e) 60–90° N, (f) 30–60° N, 120–50° W. Results are shown for the four ECLIPSE models individually and for the multi-model mean. Negative values mean that temperatures are lower in the MIT than in the CLE scenario. Error bars on the model mean values show the standard deviations of the individual model results.

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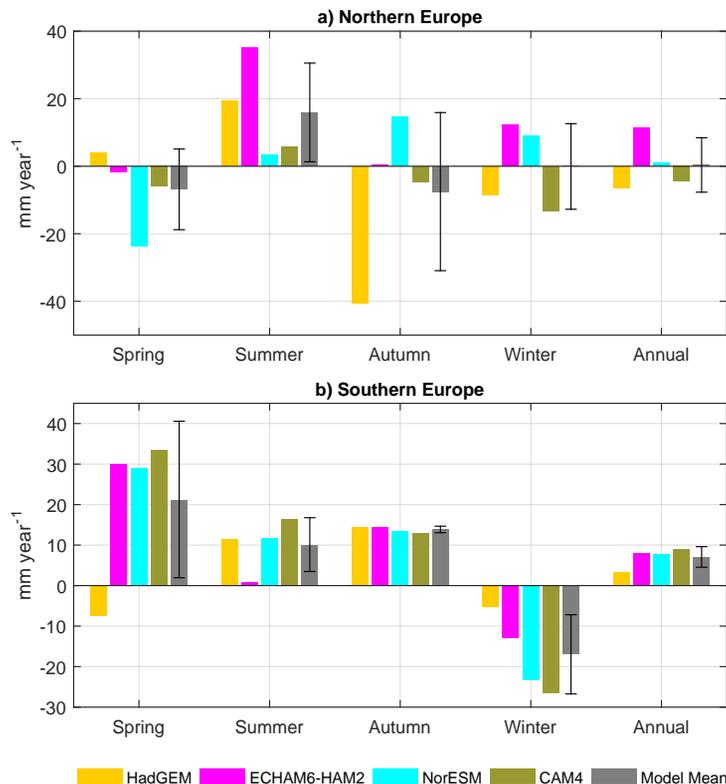


Figure 17. Seasonal and annual mean differences in precipitation (in mm yr^{-1}) over Northern (a) and Southern (b) Europe between transient simulations of the mitigation (MIT) scenario and the current legislation (CLE) scenario, averaged over the last 10 years of the simulations (2041–2050). Regions are defined as for Fig. 15a and b. Results are shown for the four ECLIPSE models individually and for the multi-model mean. Positive values mean that MIT simulations have more precipitation than CLE simulations. Error bars on the model mean values show the standard deviations of the individual model results.

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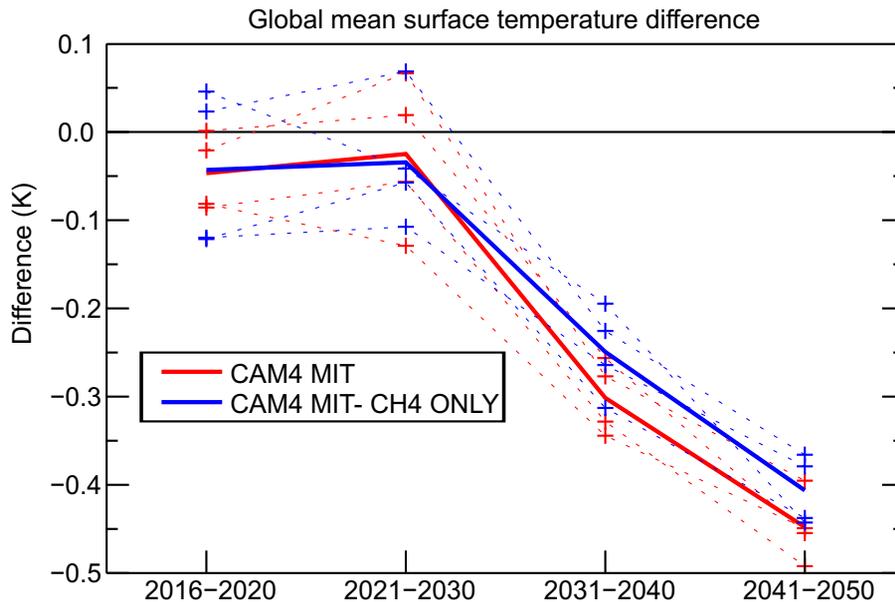


Figure 18. Time evolution of differences in global mean surface temperature between transient simulations following the mitigation (MIT) and the current legislation (CLE) scenario (i.e., MIT-CLE), as simulated by the CESM-CAM4 model with a slab-ocean representation. One experiment (red lines) included all emission reductions of the MIT scenario, whereas another experiment (blue lines) included only the CH₄ emission reductions of the MIT scenario. Ensemble mean results are shown with thick lines, individual ensemble members with thin lines.

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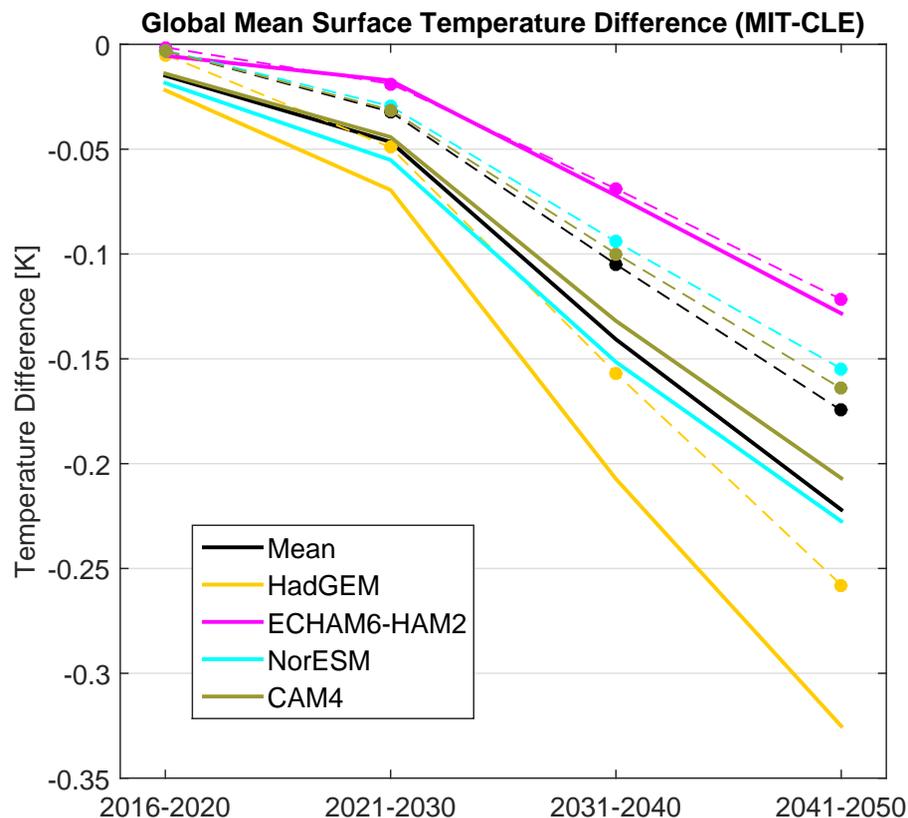


Figure 19. Time evolution of difference in global mean surface temperature for the MIT and CLE scenario calculated with the ARTP-based method. Solid lines are the total difference (MIT-CLE), while the dashed lines give the responses to CH₄ mitigation only. The solid lines can be compared directly with the results from the transient model simulations shown in Fig. 14.

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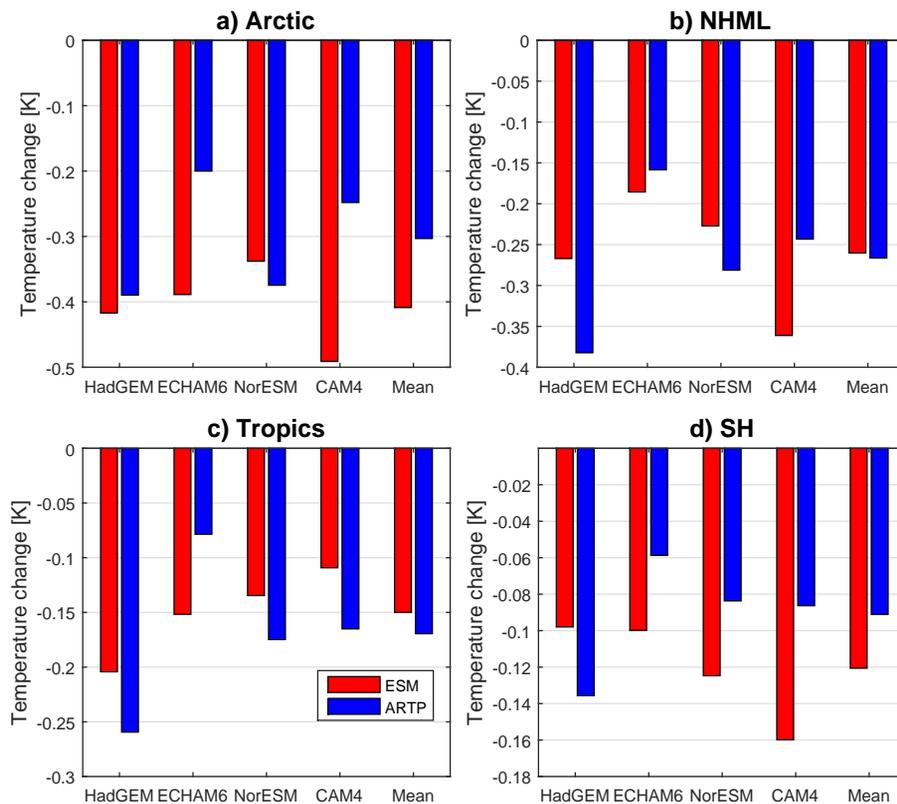


Figure 20. Temperature changes (MIT-CLE) in four latitude bands (Arctic, Northern Hemisphere Mid-Latitudes (NHML), Tropics and Southern Hemisphere) calculated with the ESMs and ARTP-based method. Mean changes over the 2031–2050 period are shown for individual models and the mean over the models. Notice the different temperature scales.

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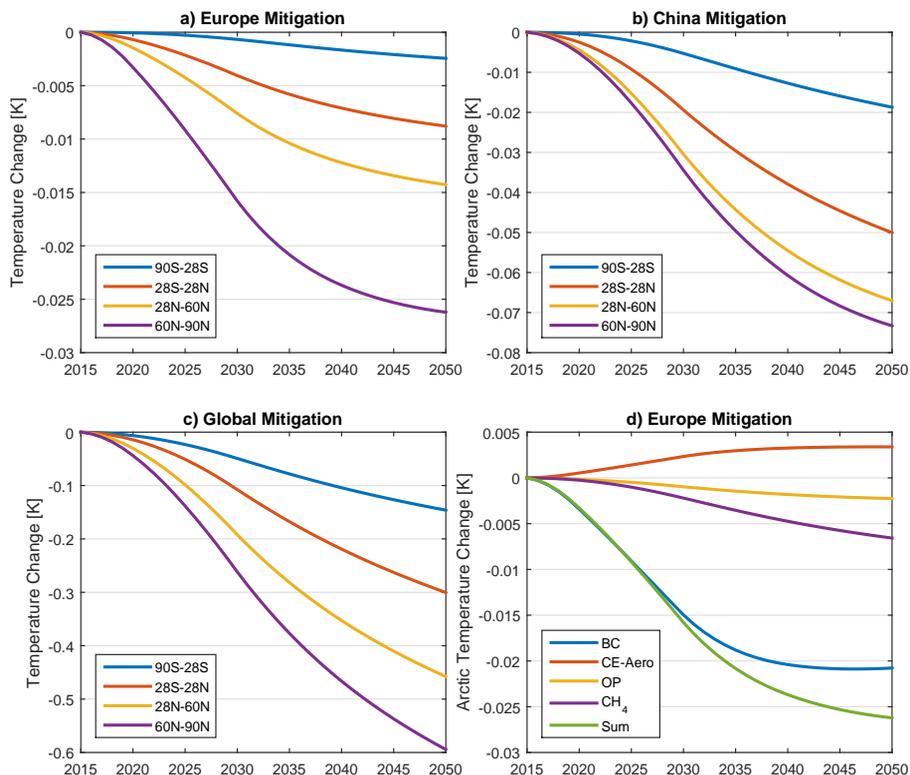


Figure 21. Annual mean surface temperature changes estimated by the ARTP method. **(a–c)** Changes in four latitude bands due to ECLIPSE mitigation scenario (MIT-CLE) for mitigation in Europe **(a)**, China **(b)** and globally **(c)**. **(d)** Arctic temperature changes due to mitigation of individual components from Europe. CE-Aero: Co-emitted aerosol (precursor) species (OA, SO₂ and NH₃), OP: Ozone precursors (NO_x, CO and NMVOCs). Note the different scales on the vertical axes.

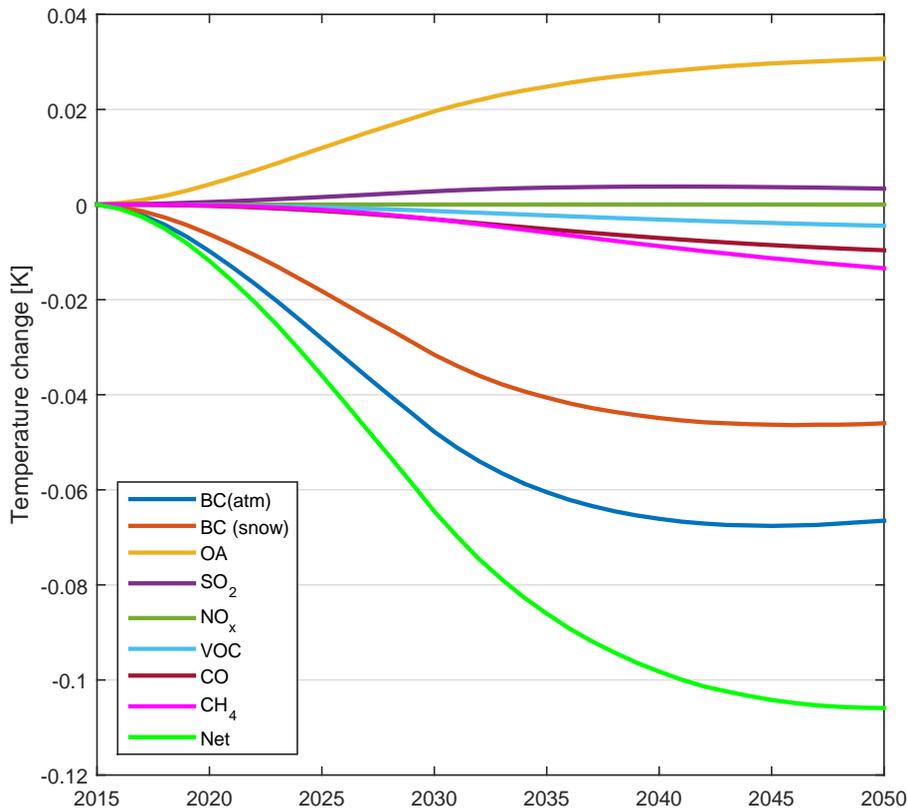


Figure 22. Total Arctic surface temperature change (K) by global mitigation of residential burning (heating and cooking) in the ECLIPSE emission scenario (MIT-CLE), as obtained with the ARTP method.

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